

# Determination of Selected Polychlorinated Biphenyls in Soil Using a OuEChERS-based Method and Gas Chromatography Tandem Mass Spectrometry 

Application Note

Environmental

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#### Abstract

Soil is an important object in pollution assessment, environmental behavior, and toxicity studies for polychlorinated biphenyls (PCBs). Accelerated solvent extraction and solid phase extraction are generally required for extraction and cleanup of PCBs in soil. These are tedious and time-consuming procedures. In this research, a modified OuEChERS procedure combined with gas chromatography-triple quadrupole mass spectrometry was developed for the determination of 20 selected PCB congeners in soil. The average recoveries from spiked soils ranged between 70 and $120 \%$, with satisfactory relative standard deviations for all the PCBs. The limits of quantitation (LOOs) were in the range of 0.01 and $0.05 \mathrm{ng} / \mathrm{g}$. The method was successfully applied to the analysis of 66 agricultural soils. The procedure proved to be simple, sensitive, efficient, and environmentally friendly.


## Introduction

Polychlorinated biphenyls (PCBs) are one of the most widespread and persistent pollutants in the world [1]. There are 209 individual PCB components, known as congeners, and, because of their persistance and hydrophobicity, they accumulate in soils where they are likely to be retained for many years. Consequently, soils are a primary reservoir for PCBs [2]. Because of their global contamination and adverse effects on environmental and human health [3], the production of PCBs has been banned worldwide since the early 1970s. Although PCBs have been banned for decades, they are still found in soils, surface waters, sediments, and air, since they are transported far from their sources [4]. Due to the increasing concern about chemical contamination of soil, there is growing interest in the scientific community and international agencies for soil pollution monitoring and assessment. To study environmental behavior in soil, it is critical to develop simple, sensitive, and reliable analytical methods.

The Quick, Easy, Cheap, Effective, Rugged, and Safe (OuEChERS) method, originally developed for extracting pesticides in fruit and vegetables in 2003 by Anastassiades et al. [5], has extended its application to many other matrices and pollutants. The QuEChERS multiresidue procedure omits or replaces many complicated analytical steps commonly used in traditional methods, making the sample preparation simple and efficient.

The aim of this research was to develop a simple, robust, and effective multiresidue method based on the OuEChERS procedure for the determination of 20 selected PCBs in soil. Different extract solvents, extract times, and cleanup adsorbents were tested and optimized. The optimum method provided a new chemical analysis method for PCB monitoring and environmental behavior studies in soil. This application note describes a recently published study of selected polychlorinated biphenyls in soil and earthworm using a QuEChERS-based method and gas chromatography tandem mass spectrometry [6]. It is a rapid multiresidue method based on a OuEChERS sample preparation, combined with GC/MS/MS detection, aiming to determine selected PCBs in soil.

## Reagents and standards

Acetonitrile, hexane, and cyclohexane were HPLC grade. Water was purified by a Milli- 0 system. PCB congeners (IUPAC numbers $28,52,77,91,95,101,105,114,118$, $136,138,149,153,157,167,169,176,180,183$, and 189 in isooctane all at $100 \mu \mathrm{~g} / \mathrm{mL}$ ), surrogate ( ${ }^{13} \mathrm{C}_{12}$-PCB 52, $40 \mu \mathrm{~g} / \mathrm{mL}$ in isooctane), and internal standard (PCB 202, $100 \mathrm{mg} / \mathrm{L}$ in isooctane) were purchased from Accustandard (New Haven, CT, USA). Stock standard solution (20 PCB mixtures) and surrogate solution of $1 \mu \mathrm{~g} / \mathrm{mL}$ were prepared in cyclohexane, respectively. A $0.1 \mu \mathrm{~g} / \mathrm{mL}$ internal standard solution was prepared in cyclohexane. Ceramic homogenizers and dispersive solid phase extraction adsorbent (PSA, C18, $\mathrm{MgSO}_{4}$ ) were purchased from Agilent Technologies (CA, USA). Certified reference material of seven PCBs in soil was obtained from Agro-Environmental Quality Supervision \& Testing Center, MOA (Tianjin, China).

## Instrument conditions

GC Conditions

| GC system | Agilent 7890A, coupled with an Agilent 7693 autosampler |
| :---: | :---: |
| Column | Agilent HP-5 MS UI ( $30 \mathrm{~m} \times 0.25 \mathrm{~mm}, 0.25 \mu \mathrm{~m}$ ) ( $\mathrm{p} / \mathrm{n}$ 19091S-433 UI) |
| Oven temperature | $60^{\circ} \mathrm{C}$ hold 1 minute, at $40^{\circ} \mathrm{C} / \mathrm{min}$ to $120^{\circ} \mathrm{C}$, at $5^{\circ} \mathrm{C} / \mathrm{min}$ to $275^{\circ} \mathrm{C}$ |
| Carrier gas | Helium |
| Flow rate | $1.0 \mathrm{~mL} / \mathrm{min}$ |
| Injection port temperature | $280{ }^{\circ} \mathrm{C}$ |
| Injection volume | $1.0 \mu \mathrm{~L}$ |
| Injection mode | Splitless, purge on after 1.5 minutes |
| MS Conditions |  |
| MS system | Agilent 7000C Triple Quadrupole GC/MS System |
| Ion source | El |
| Ionization voltage | 70 eV |
| Ion source temperature | $280{ }^{\circ} \mathrm{C}$ |
| Quadrupole temperature | $\begin{aligned} & 01150^{\circ} \mathrm{C} \\ & 02150^{\circ} \mathrm{C} \end{aligned}$ |
| Interface temperature | $280{ }^{\circ} \mathrm{C}$ |
| Solvent delay | 10.0 minutes |

Table 1 gives the retention time and MRM transition parameters for the selected PCB congeners.

## Sample preparation

Soil used in the recovery experiment was collected from an agriculture field in Tianjin, China. The soil was air-dried at room temperature, ground, and sieved through a $2-\mathrm{mm}$ mesh. The physicochemical properties of the soil were as follows: pH 7.53, 1.53 \% organic matter, 16.72 \% clay, 43.25 \% sand, and 39.71 \% silt.

Table 1. Retention Time and MRM Transition Parameters for the Selected PCB Congeners

| Compound | $\mathbf{t}_{\mathbf{R}}(\mathbf{m i n})$ | MRM1 | CE | MRM2 | CE | MRM3 | CE |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| PCB28 | 17.75 | $256 \rightarrow 186$ | 25 | $258 \rightarrow 188$ | 25 | $258 \rightarrow 186$ | 25 |
| ${ }^{13}$ C $_{12}$-PCB52a | 19.05 | $232 \rightarrow 162$ | 40 | $232 \rightarrow 197$ | 30 | $304 \rightarrow 269$ | 15 |
| PCB52 | 19.06 | $290 \rightarrow 220$ | 25 | $292 \rightarrow 222$ | 25 | $292 \rightarrow 220$ | 25 |
| PCB95 | 21.46 | $326 \rightarrow 256$ | 25 | $328 \rightarrow 256$ | 30 | $326 \rightarrow 291$ | 15 |
| PCB91 | 21.69 | $326 \rightarrow 256$ | 25 | $328 \rightarrow 256$ | 25 | $326 \rightarrow 291$ | 15 |
| PCB101 | 22.31 | $326 \rightarrow 256$ | 30 | $328 \rightarrow 256$ | 30 | $326 \rightarrow 254$ | 30 |
| PCB136 | 23.53 | $360 \rightarrow 290$ | 30 | $362 \rightarrow 290$ | 30 | $360 \rightarrow 288$ | 30 |
| PCB77 | 23.63 | $290 \rightarrow 220$ | 25 | $292 \rightarrow 222$ | 25 | $220 \rightarrow 150$ | 35 |
| PCB149 | 24.51 | $360 \rightarrow 290$ | 30 | $362 \rightarrow 290$ | 30 | $360 \rightarrow 325$ | 15 |
| PCB118 | 24.57 | $326 \rightarrow 256$ | 30 | $328 \rightarrow 256$ | 30 | $326 \rightarrow 254$ | 30 |
| PCB114 | 24.97 | $326 \rightarrow 256$ | 25 | $328 \rightarrow 256$ | 25 | $326 \rightarrow 254$ | 25 |
| PCB153 | 25.39 | $360 \rightarrow 290$ | 25 | $362 \rightarrow 290$ | 25 | $360 \rightarrow 288$ | 25 |
| PCB105 | 25.52 | $326 \rightarrow 256$ | 30 | $328 \rightarrow 256$ | 30 | $326 \rightarrow 254$ | 30 |
| PCB176 | 26.16 | $394 \rightarrow 324$ | 25 | $396 \rightarrow 324$ | 25 | $396 \rightarrow 326$ | 25 |
| PCB138 | 26.37 | $360 \rightarrow 290$ | 30 | $360 \rightarrow 325$ | 15 | $362 \rightarrow 290$ | 30 |
| PCB183 | 27.14 | $394 \rightarrow 324$ | 30 | $396 \rightarrow 326$ | 30 | $396 \rightarrow 324$ | 30 |
| PCB167 | 27.36 | $360 \rightarrow 290$ | 25 | $362 \rightarrow 290$ | 25 | $360 \rightarrow 288$ | 30 |
| PCB202 | 28.10 | $428 \rightarrow 358$ | 30 | $430 \rightarrow 360$ | 30 | $430 \rightarrow 358$ | 30 |
| PCB157 | 28.35 | $360 \rightarrow 290$ | 30 | $362 \rightarrow 290$ | 30 | $360 \rightarrow 288$ | 30 |
| PCB180 | 28.74 | $394 \rightarrow 324$ | 30 | $394 \rightarrow 359$ | 15 | $396 \rightarrow 326$ | 30 |
| PCB169 | 29.45 | $360 \rightarrow 290$ | 30 | $362 \rightarrow 290$ | 30 | $360 \rightarrow 288$ | 30 |
| PCB189 | 30.72 | $396 \rightarrow 324$ | 30 | $396 \rightarrow 326$ | 30 | $394 \rightarrow 324$ | 30 |

${ }^{\text {a }}$ surrogate standard
${ }^{\mathrm{b}}$ internal standard

Optimized extraction and cleanup procedure

*The Dispersive solid phase extraction adsorbents kits for soil ( $\mathrm{p} / \mathrm{n} 5982-5156$ )

## Results and Discussion

## Method validation

The analytical parameters including linearity, limit of quantitation (LOO), repeatability, and recoveries were studied under the optimized extraction and cleanup conditions. To determine the method accuracy, a recovery study was carried out, comparing the concentration of each PCB measured by performing the complete procedure with the known concentration fortified to blank soil samples at 0.1, 1 , and $10 \mathrm{ng} / \mathrm{g}$ in replicates $(\mathrm{n}=5)$. Blank soil samples were preanalyzed to verify the absence of selected PCBs. The matrix-dependent LOO of the method were determined using spiked blank samples, and defined as the lowest fortified concentrations that produce a signal-to-noise (S/N) ratio above 10. Calibration curves were calculated with standards in solvent at concentrations of $0.1,1,5,10$, and $50 \mathrm{ng} / \mathrm{mL}$ for each PCB.

Good linearity was obtained for all the PCB congeners within the concentration range of 0.1 to $50 \mathrm{ng} / \mathrm{mL}$, with a coefficient of detection $\left(R^{2}\right)$ higher than 0.999 . Table 2 shows the results of the recovery experiments for soil under optimized extraction and cleanup conditions.

Table 2. LOOs, Recoveries, and Relative Standard Deviation (RSDs) of Selected PCBs for Soil Spiked at $0.1 \mathrm{ng} / \mathrm{g}, 1 \mathrm{ng} / \mathrm{g}$, and $10 \mathrm{ng} / \mathrm{g}$

|  | 0.1 (ng/g) |  | 1 (ng/g) |  | 10 ( $\mathrm{ng} / \mathrm{g}$ ) |  | $\begin{aligned} & \mathrm{LOO} \\ & (\mathrm{ng} / \mathrm{g}) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Rec (\%) | RSD (\%) | Rec (\%) | RSD (\%) | Rec (\%) | RSD (\%) |  |
| PCB28 | 112.9 | 20.1 | 108.0 | 2.0 | 102.6 | 5.7 | 0.02 |
| ${ }^{13} \mathrm{C}_{12}$-PCB52 | 105.1 | 3.7 | 98.7 | 2.4 | 105.5 | 3.2 | 0.01 |
| PCB52 | 115.0 | 9.0 | 105.7 | 1.8 | 107.8 | 3.9 | 0.01 |
| PCB95 | 108.9 | 7.9 | 100.9 | 1.4 | 109.6 | 2.7 | 0.01 |
| PCB91 | 108.9 | 6.0 | 98.5 | 2.3 | 107.7 | 3.3 | 0.02 |
| PCB101 | 116.5 | 6.7 | 96.9 | 1.7 | 102.3 | 3.8 | 0.02 |
| PCB136 | 114.9 | 5.2 | 99.4 | 0.4 | 108.9 | 2.4 | 0.01 |
| PCB77 | 94.8 | 8.7 | 93.7 | 1.4 | 100.7 | 5.5 | 0.02 |
| PCB149 | 105.8 | 10.2 | 98.1 | 0.9 | 100.1 | 4.2 | 0.01 |
| PCB118 | 95.0 | 7.3 | 90.6 | 0.8 | 95.3 | 4.6 | 0.02 |
| PCB114 | 93.3 | 3.4 | 88.8 | 1.5 | 95.9 | 4.5 | 0.02 |
| PCB153 | 99.8 | 1.9 | 83.2 | 3.0 | 87.1 | 5.0 | 0.02 |
| PCB105 | 115.5 | 9.8 | 91.0 | 5.0 | 99.9 | 3.6 | 0.02 |
| PCB176 | 110.9 | 11.5 | 85.0 | 2.8 | 94.7 | 5.0 | 0.01 |
| PCB138 | 117.7 | 6.1 | 90.1 | 5.0 | 96.1 | 3.1 | 0.01 |
| PCB183 | 96.7 | 13.6 | 79.3 | 2.2 | 84.1 | 6.6 | 0.02 |
| PCB167 | 80.3 | 10.7 | 80.8 | 5.5 | 84.6 | 4.4 | 0.02 |
| PCB157 | 98.9 | 8.7 | 84.5 | 2.1 | 89.2 | 5.1 | 0.05 |
| PCB180 | 89.5 | 15.8 | 79.0 | 2.6 | 83.0 | 5.8 | 0.05 |
| PCB169 | 92.2 | 12.2 | 72.4 | 3.5 | 70.1 | 6.2 | 0.05 |
| PCB189 | 83.6 | 13.1 | 70.0 | 2.5 | 76.2 | 5.3 | 0.02 |

All recoveries at $0.1,1$, and $10 \mathrm{ng} / \mathrm{g}$ ranged between 70 and $120 \%$, with the majority of recoveries greater than $90 \%$ at all concentrations. The precision was satisfactory, with the majority of RSDs below $10 \%$, which complies with SANCO/12471/2013 [7]. In general, the PCBs with seven chlorine atoms had relatively lower recoveries than those with six or fewer chlorine atoms. The LOOs of these 20 PCB congeners in soil ranged between 0.01 and $0.05 \mathrm{ng} / \mathrm{g}$, which were lower than those previously reported by other authors [8-10]. Table 2 gives the resulting LOOs.

As described above, the performance of the present method was satisfactory, with low LOOs as well as good recovery and precision. Compared with traditional methods such as soxhlet extraction and accelerated solvent extraction, the proposed method used less organic solvent (only 10 mL of acetonitrile was needed), a shorter extraction time ( 3 minutes of vortexing), and no special equipment, which made the method simple and environmentally friendly.

## Analysis of certified reference material

Metrological traceability is essential to ensure that measurement results are comparable in time and space [11]. The use of certified reference material (CRM) provides quality assurance in environmental analysis, and makes reliable and tractable analytical results possible. To further test the QuEChERS-based method, we analyzed CRM BW 3714, seven PCBs in soil. The results for the PCB-containing CRM is listed in Table 3. The results obtained by the OuEChERS-based method were in good agreement with the certified value of the CRM. All the results were within the uncertainty range given in the CRM, and good repeatability was obtained.

Table 3. Results from the Analysis of CRM Using the QuEChERS-based Method

| PCBs | BW3714 |  | Analysis results |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Certified value ( $\mathrm{ng} / \mathrm{g}$ ) | Uncertainty range ( $\mathrm{ng} / \mathrm{g}$ ) | Analyzed value ( $\mathrm{ng} / \mathrm{g}$ ) | $\begin{aligned} & \text { SD } \\ & (\mathrm{n}=3) \end{aligned}$ |
| PCB28 | 7.36 | 6.63-8.09 | 6.80 | 0.03 |
| PCB52 | 1.40 | 1.23-1.57 | 1.33 | 0.04 |
| PCB101 | 1.45 | 1.32-1.58 | 1.57 | 0.07 |
| PCB118 | 4.56 | 4.20-4.92 | 4.79 | 0.10 |
| PCB153 | 0.64 | 0.60-0.68 | 0.64 | 0.04 |
| PCB138 | 0.89 | 0.78-1.00 | 0.81 | 0.04 |
| PCB180 | 0.23 | 0.18-0.28 | 0.23 | 0.03 |

## Analysis of real samples

The method was used to analyze 66 soil samples collected from agricultural fields located in different cities in Shandong Province, China. Soil samples 1-16 were collected from rice fields, vineyards, and apple orchards. Samples 17-66 were collected in greenhouses from Shouguang, the largest vegetable base in China, located in east Shandong province. Soil was sampled from the upper layer ( $0-30 \mathrm{~cm}$ ), and transported to the laboratory where it was air-dried and sieved through a $2-\mathrm{mm}$ mesh. The soil samples were weighed into a $50-\mathrm{mL}$ centrifuge tube, and 0.5 mL of surrogate $(25 \mathrm{ng} / \mathrm{mL})$ was added. The samples were allowed to stand for 24 hours in room temperature before extraction. Nine of the 20 PCB congeners were identified in 16 soil samples, with four of them being marker PCBs (PCB 28, 52, 138, and 153). The detection rates and concentrations of the PCBs were typically low. Only seven samples were found to contain PCB congeners with concentrations higher than the corresponding LOQs. Most of the detected PCBs were at concentrations lower than the LOOs. Compared to the low concentrations of PCBs in agricultural fields, the detection rate and concentrations in the E-Waste site soil were much higher [12-13].

## Conclusions

This application note demonstrates the applicability of the QuEChERS-based procedure, combined with GC/MS/MS for the determination of 20 selected PCBs in soil. Satisfactory validation parameters including linearity, LOO, and RSD were obtained, demonstrating the feasibility of the method. The method was applied for 66 agricultural soils, and nine of the 20 PCBs were detected at low concentrations. This simple and sensitive method is expected to provide a new chemical analysis method for PCB monitoring and environmental behavior studies in soil.

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